Nearshore seascape connectivity enhances seagrass meadow nursery function

Angeleen M. Olson,^{1,2,4} Margot Hessing-Lewis,² Dana Haggarty,³ and Francis Juanes¹

¹Department of Biology, University of Victoria, 3800 Finnerty Road, Victoria, British Columbia V8P 5C2 Canada ²Hakai Institute, P.O. Box 309, Heriot Bay, British Columbia V0P 1H0 Canada ³Stock Assessment and Research Division, Pacific Biological Station, Fisheries and Oceans Canada, 3190 Hammond Bay Road, Nanaimo, British Columbia V9T 6N7 Canada

Citation: Olson, A. M., M. Hessing-Lewis, D. Haggarty, and F. Juanes. 2019. Nearshore seascape connectivity enhances seagrass meadow nursery function. Ecological Applications 29(5): e01897. 10.1002/eap.1897

Abstract. Diverse habitats composing coastal seascapes occur in close proximity, connected by the flux of materials and fauna across habitat boundaries. Understanding how seascape connectivity alters important ecosystem functions for fish, however, is not well established. For a seagrass-dominant seascape, we predicted that configuration and composition of adjacent habitats would alter habitat access for fauna and trophic subsidies, enhancing nursery function for juvenile fish. In an extensive Zostera marina seagrass meadow, we established sites adjacent to (1) highly complex and productive kelp forests (Nereocystis luetkeana), (2) unvegetated sand habitats, and (3) in the seagrass meadow interior. Using SCUBA, we conducted underwater observations of young-of the-year (YOY) rockfish (Sebastes spp.) recruitment across sites. Using generalized linear mixed effects models, we assessed the role of seascape adjacency relative to seagrass provisions (habitat complexity and prey) on YOY recruitment. YOY rockfish collections were used to trace sources of allochthonous vs. autochthonous primary production in the seagrass food web, via a δ^{13} C and δ^{15} N isotopic mixing model, and prey consumption using stomach contents. Overall, seagrass nursery function was strongly influenced by adjacent habitats and associated subsidies. Allochthonous N. luetkeana was the greatest source of energy assimilated by YOY rockfish within seagrass sites. In seagrass sites adjacent to N. luetkeana kelp forests, YOYs consumed higher quality prey, which corresponded with better body condition relative to sites adjacent to sand. Moreover, kelp forest adjacency enhanced YOY rockfish recruitment within the seagrass meadow, suggesting that habitat complexity is a key seascape feature influencing the nursery function of nearshore habitats. In general, to promote seascape connectivity, the conservation and restoration of nursery habitats should prioritize the inclusion of habitat mosaics of high structural complexity and productivity. We illustrate and emphasize the importance of using a seascape-level approach that considers linkages among habitats for the management of important nearshore ecosystem functions.

Key words: eelgrass; food web; habitat complexity; habitat configuration; kelp forests; predator-prey interactions; rockfish; spatial subsidies; stable isotopes; Zostera marina.

INTRODUCTION

Ecosystems are naturally linked by fluxes across habitat boundaries, influencing the structure and function of communities (Loreau et al. 2003). However, accelerating habitat loss and fragmentation from anthropogenic activities are altering landscape and seascape connectivity worldwide (Haddad et al. 2015). Accordingly, declines in habitat-dependent biota and key ecosystem functions are widespread (Ellison et al. 2005, Hughes et al. 2009). Yet, in the oceans, little is known about how

Manuscript received 26 September 2018; revised 25 February 2019; accepted 19 March 2019. Corresponding Editor: Paul K. Dayton.

⁴E-mail: angeleen.olson@hakai.org

seascape connectivity affects nearshore ecosystem functions, such as nursery effects, and how such functions will respond to the progression of habitat loss (Boström et al. 2011), or have the potential to mitigate this loss (Olds et al. 2015).

"Seascapes" are a spatially defined marine area consisting of multiple habitat patches. Seascapes are connected by proximity to neighboring habitats and the movement of fauna, material, and nutrients, herein "seascape connectivity" (Grober-Dunsmore et al. 2009, Sheaves 2009). Seascape connectivity may alter the role of habitats in supporting marine communities when linkages among habitats are established and/or lost (Ries et al. 2004, Ellison et al. 2005, Hughes et al. 2009, Thompson et al. 2017). Within seascapes, specific habitat attributes ("seascape features") can mediate the flow and access of materials among habitats, resulting in varying effects on biological communities. These interactive seascape features can be grouped into broad categories: (1) the spatial arrangement of habitats and (2) the composition of habitat (Olds et al. 2015, Staveley et al. 2017, Van Wynsberge et al. 2017).

The spatial arrangement of habitats can alter movement patterns of fauna in the seascape via edge effects, corridors, and patch sizes, by providing optimal areas for shelter, foraging, or migration (McMahon et al. 2012, Berkström et al. 2013). Edges, distinct boundaries between two habitats, can facilitate access to spatially distinct subsidies, such as nutrients and prey, or access to habitat (Fagan et al. 1999, Ries et al. 2004, Rand et al. 2006). In addition, the composition (i.e., type) of habitats can further alter connectivity depending on their structural complexity. Habitats with high structural complexity (e.g., mangroves, biogenic reefs) offer contrasting shelter, productivity, and provision of subsidies compared to unvegetated habitats with low structural complexity (e.g., unvegetated sand). Mediated by these seascape features and transported by hydrodynamics forces or mobile fauna, resources from donor habitats can dramatically affect consumer-resource dynamics that enhance community diversity and/or densities in recipient habitats (i.e., "spatial subsidies"; Polis et al. 1997, Fagan et al. 1999, Massol et al. 2011); important components of nursery function (Fig. 1).

Located within nearshore habitat mosaics spanning coastlines globally, seagrass meadows are an important nursery habitat for many marine fishes and invertebrates (Beck et al. 2001, Nagelkerken et al. 2002, Dorenbosch et al. 2005). Seagrass meadows provide complexity in the water column, increase faunal refugia, support epiphytic algal production, and in turn, sustain invertebrate grazers available for consumers (Orth et al. 1984, Kennedy et al. 2018). Seagrass meadows, therefore, tend to have higher juvenile densities, survival, and growth, potentially resulting in increased movements to adult habitats relative to unvegetated habitats (Heck et al. 2003, McDevitt-Irwin et al. 2016). Seagrass meadow nurseries have primarily been studied at a single-habitat scale, but it is increasingly acknowledged that they are highly connected to their surrounding habitats (McMahon et al. 2012, Olds et al. 2012, Hyndes et al. 2014). Recent evidence suggests that fluxes of nutrients and materials (Davis et al. 2014, Ricart et al. 2015), as well as fauna (Staveley et al. 2017, Perry et al. 2018) to seagrass meadows are spatially mediated by adjacent habitats. As seagrass and other nearshore habitats decline worldwide (Orth et al. 2006, Waycott et al. 2009), understanding how seagrass nursery function is affected by seascape connectivity can fill a critical knowledge gap for coastal conservation and management that often occurs at seascape-level scales (Boström et al. 2006, Nagelkerken et al. 2013).

In marine environments, the cross-boundary exchange of resources between ecosystems (Dale and

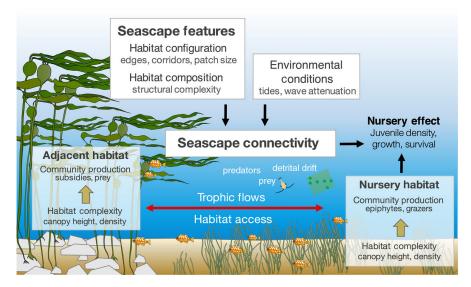


FIG. 1. A conceptual diagram of seascape connectivity influencing nearshore nursery habitats. Seascape features, such as (1) habitat configuration and (2) habitat composition, together with environmental processes, can mediate trophic flows and proximity between nearshore habitats (i.e., seascape connectivity). Nursery habitats and their adjacent habitat neighbors vary in their structural complexity, which influences their overall community productivity (e.g., basal production, prey diversity, epiphyte biomass), as well vary in spatial configuration to one another. As a result, key seascape features may facilitate connections between habitats, via allochthonous subsidies (prey and detritus) or habitat access, and in turn, influence juvenile densities, growth, and survival in nursery habitats.

Prego 2002, Naiman et al. 2002), and between tropical seascape habitats (Berkström et al. 2013, Davis et al. 2014) is well documented. Yet, empirical studies on connectivity between high complexity temperate coastal habitats are limited (McDevitt-Irwin et al. 2016). Kelp forests and seagrass habitats can form coastal habitat mosaics along northern temperate coastlines, where kelp forests may support seagrass nurseries in a number of ways (Fig. 1). Adding vertical structure and biomass to the water column, kelp forests increase habitat complexity available for organisms (Graham 2004, Trebilco et al. 2015), as well as export nutrients and detritus into the marine environment (Krumhansl and Scheibling 2012, Hyndes et al. 2014). This allochthonous production can become important food sources to recipient consumers in seagrass meadows (Doropoulos et al. 2009, Hyndes et al. 2012), and has been observed to promote secondary production in other marine systems (Kelly et al. 2012).

In the northeast Pacific, temperate seagrass meadows are nurseries for rockfishes (Sebastes spp.): a biologically, culturally, and commercially important group of fishes in decline due to overfishing (Love et al. 2002, Yamanaka and Logan 2010, Eckert et al. 2017). After a pelagic larval stage, some rockfish species recruit to shallow nearshore habitats, such as seagrass meadows and kelp forests, before transitioning to deeper adult habitats (Love et al. 1991). Post-recruitment processes that can be mediated by habitat complexity, such as reduced mortality or increased growth, are critical during these juvenile stages (Johnson 2007, Juanes 2007) and may regulate population success (Haggarty et al. 2017, Tupper and Juanes 2017). Thus, increased habitat access and subsidies from seascape connections may affect juvenile rockfishes by increasing their growth or survival (Kamimura and Shoji 2013).

Determining how connectivity between seagrass meadows and kelp forests influences the nursery function for declining rockfish populations is important to nearshore habitat management. In response, this study investigated how adjacent habitats in a seagrass-kelpforest-sand seascape influenced components of seagrass meadow nursery function. We observed young-of-theyear (YOY) rockfish recruitment and feeding ecology within a large temperate Zostera marina meadow at seagrass sites adjacent to kelp forests (Nereocystis luetkeana), unvegetated sand habitats, and in the meadow interior (Fig. 2). We hypothesized that adjacent N. luetkeana kelp forests would enhance seagrass meadow nursery function through two mechanisms: (1) increased allochthonous subsidies (primary production and prey) to YOY rockfish diets and (2) greater YOY rockfish recruitment to the meadow edge where habitat complexity is maximized from both seagrass and kelp forest three-dimensional structure (Connell and Jones 1991). Conversely, we predicted that nursery function would decrease at sites adjacent to unvegetated sand due to a lack of allochthonous inputs and habitat complexity.

Methods

Study area and design

We conducted this study in a subtidal Z. marina meadow in Choked Passage on British Columbia's (Canada) central coast (Fig. 2a, b). The shoreline complexity of the central coast allows for a diverse matrix of nearshore marine habitats within this seascape (Hessing-Lewis et al. 2018). Similar to other northern-latitude coastlines, mixed patches of seagrass, kelp forest habitats (N. luetkeana and/or Macrocystis pyrifera), and sand habitats are common elements of this region. The expansive (367,300 m²) and near-contiguous seagrass meadow is situated between a large outer coast island (Calvert Island) and exposed rocky islets to the west. Canopyforming N. luetkeana kelp beds propagate from the adjacent rocky reefs, while unvegetated sand flats border Calvert Island's shoreline and deeper waters (Fig. 2c). The Z. marina meadow is, on average, 3.5 m deep and is subject to high-current tidal exchanges.

We used SCUBA surveys to identify seascape features around the Z. marina meadow, based on the arrangement (herein, "edge"), and composition of adjacent habitat: high complexity (N. luetkeana kelp forest) or low complexity (unvegetated sand). Once mapped, we partitioned the perimeter into 50-m sections, categorized the sections by adjacent habitat type, and assigned them a unique number, which was used for random site selection. We randomly selected seagrass edge sections (n = 4 per edge type) as study sites (Fig. 2c, white circles and squares, kelp edge and sand edge, respectively). The interior of the seagrass meadow (>200 m from edges) was also mapped, and randomized coordinates and compass directions were used to establish seagrass interior sites (n = 4, Fig. 2c, trian-)gles). As rockfish have high site-fidelity during early life stages (Hoelzer 1988, Nelson 2001), we assumed biological site independence by maintaining a minimum distance of 100 m between meadow sites during siteselection. For additional habitat comparisons of recruitment, sites were also established within the N. luetkeana kelp forests (n = 4) and sand habitats (n = 4), directly adjacent to the seagrass edge sites (Fig. 2c, gray symbols).

Food web field collections and analyses

To quantify seascape subsidies in the seagrass meadow, we collected key constituents of the seagrass food web: YOY rockfish, invertebrate prey, and primary producers.

YOY Rockfish—In late August to early September 2015, we used a modified otter trawl (10-mm mesh) to capture

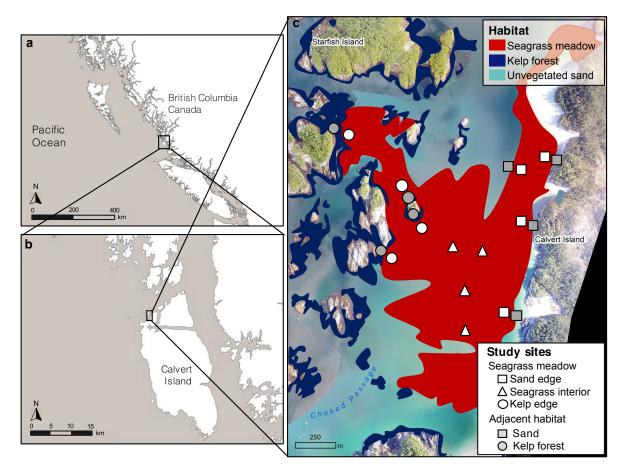


FIG. 2. (a, b) Map of the study area located northwest of Calvert Island, British Columbia, Canada and (c) the focal nearshore seascape in Choked Passage. The seagrass meadow (red) is ~367,000 m², bordered by *Nereocystis luetkeana* kelp forests (dark blue) and unvegetated sand (light blue). Study sites in the seagrass meadow (white) include seagrass interior (triangles), *N. luetkeana* kelp edge (circles, high complexity adjacent habitat), and sand edge sites (squares, low complexity adjacent habitat). Sites were also established within the adjacent habitats: *N. luetkeana* kelp forest (gray circles) and unvegetated sand sites (gray squares).

recently recruited YOY rockfish at seagrass sites. The trawl was towed from a small boat at slow speed for 5 min through the seagrass canopy. The majority of rockfish captured were identified as part of the copper rockfish (*Sebastes caurinus*)–quillback rockfish (*Sebastes maliger*) juvenile complex. YOY copper-quillbacks were immediately sacrificed by blunt force, and measured for total length (mm) and mass (g). Stomachs were dissected and stored in ethanol for dietary analysis. The remaining whole bodies were frozen for isotope analysis.

Stomach content analysis—Prey items retrieved from the stomachs were identified to the lowest taxonomic level possible using a dissecting microscope and grouped by similar taxonomy (Appendix S1: Table S1). Prey was counted and weighed to the nearest milligram per prey group. A conservative count of 1 was assigned to a group if digestion affected the intactness of prey. If a group was too light to register a mass, a conservative 0.001 mg was assigned. Proportional representation of a prey item, based on the total sum of prey across each site type (e.g., sand edge, seagrass interior, and kelp edge), was estimated using percent numeric (% N), gravimetric (% G), and frequency of occurrence (% O). An index of relative importance (% IRI) facilitated comparisons (Liao et al. 2001, Hart et al. 2002), and was used in the calculations of prey diversity (Shannon index; H') and Pielou's evenness (J), defined as the similarity of prey items in consumer diet (Appendix S1: Section S1).

Primary producers—Dominant primary producers were haphazardly collected from all sites by SCUBA to estimate source contributions to the food web in an isotopic mixing model. Seagrass-associated sources (i.e., autochthonous) collected were *Z. marina*, and its locally dominant epiphytic algae, *Smithora naiadum*. Allochthonous sources were *N. luetkeana* blades collected from individual plants in the adjacent kelp forest sites. Other common macroalgae, *Cymathaere triplicata* and *Alaria marginata* (herein "understory kelps"), were collected because of their presence in small patches or as drift in or near the seagrass meadow. Particulate organic matter (POM) was collected using plankton tows at 0 and 5 m depths over the seagrass meadow and 2 km outside of the seagrass bed, to represent oceanic conditions. Samples were stored frozen until laboratory processing. YOY rockfish and primary producers were collected within the same month to minimize confounding temporal changes in their isotopic signatures.

Isotope preparation and analysis—Primary producers were defrosted and cleaned of contaminating materials using visual inspection and gentle scraping. The longest Z. marina blade per shoot and the total epiphyte biomass per shoot were used for isotopic analysis. For N. luetkeana, C. triplicata, and A. marginata, a subsample of tissue from the outermost section of the blade was used for analysis. Due to the small body sizes of the YOY rockfishes, the whole body (except the gut) was used for isotopic analysis. All samples were rinsed in two baths of deionized water. The samples were dried at 60°C for 48 h and packed in tin capsules for isotopic measurement using a Delta IV Isotope Ratio Mass Spectrometer at the University of Victoria. $\delta^{15}N$ and $\delta^{13}C$ signatures were calculated using the ratio of heavy to light isotope relative to a laboratory standard and were measured in per mil (%):

$$\delta^{15}$$
N/ δ^{13} C = [($R_{\text{sample}}/R_{\text{standard}}) - 1$] × 10³

where $R = {}^{15}N{}^{14}N$ or ${}^{13}C{}^{12}C$, respectively. Carbon to nitrogen (C:N) ratios were calculated using the mass ratio between the %C and %N in the samples. Because lipids contain mostly carbon, and little nitrogen, it is assumed that an increase in C:N value positively correlates with body condition. For juvenile rockfish, lipids may be a more accurate indicator of nutritional condition than inferences by length and mass regressions (Norton et al. 2001). Analysis of variance (ANOVA) and Tukey's HSD were used to test for differences in rockfish body condition (C:N ratio) among seagrass sites.

Mixing model-Using stable isotopes, we applied a Bayesian mixing model to estimate the relative contributions of autochthonous vs. allochthonous basal energy to YOY rockfish collected in seagrass sites: sand edge (n = 25), seagrass interior (n = 30), and kelp edge (n = 33; Appendix S2: Table S1). To trace energy movement from primary producer to consumer, trophic enrichment (TE) of consumer signatures needed to be accounted for. TE is the process in which the heavy isotope is preferentially assimilated into consumer tissue via metabolic processes, which results in an enrichment bias in consumer isotopic signatures (Deniro and Epstein 1981, Minagawa and Wada 1984). Discrimination factors (change per mil, $\Delta \%$) are used to estimate TE in consumer-prey interactions. By deducting TE from consumer isotope signatures based on prey interactions, it is possible to trace backwards to the isotopic signatures of primary producers from which energy was derived (Reid et al. 2008, Phillips et al. 2014).

To do this, we constructed a seagrass trophic model to estimate predatory-prey interactions in this seagrass food web (Appendix S2: Section S1, Fig. S1) based on long-term monitoring of the Choked Passage seagrass food web (Hessing-Lewis et al. 2018) and established trophic relationships in temperate seagrass food webs (Appendix S1, S2; Hughes et al. 2013). The seagrass trophic model outlines the consumption of primary sources (e.g., seagrass, macroalgae) by invertebrate grazers, which in turn, are consumed by YOY rockfish. Discrimination factors were assigned to each trophic interaction using averaged values from McCutchan et al. (2003), based on our isotope sample processing methods and consumer characteristics. Total TE was summed across all trophic levels and deducted from YOY rockfish signatures (Appendix S2: Fig. S2), after the latter were normalized for lipids as per Post et al. (2007). Mixing models are highly sensitive to discrimination factors (Bond and Diamond 2011), thus we incorporated estimates of uncertainty in TE (Appendix S2: Fig. S1) and assessed TE-corrected consumer signatures to be within an acceptable range of primary producer signatures (Appendix S2: Fig. S2; Phillips et al. 2014).

We used the R package MixSIAR (Version 3.1.10; Moore and Semmens 2008) to estimate the relative proportions of Z. marina (n = 25), S. naiadum (n = 17), N. luetkeana (n = 5), understory kelp (n = 10), and POM (n = 8) contributing to YOY rockfish. C. triplicata and A. marginata were grouped as understory kelps according to isotopic similarity, relative palatability, and biological characteristics (Phillips et al. 2014). An uninformative prior was used to give an equal probability of consumption among primary producers and site type was analyzed as a fixed effect. To retrieve the posterior density estimates of source contribution, model iterations were run 10⁵ times, with the first half of the estimations discarded.

Observational surveys

To assess rockfish recruitment across the seascape, we used SCUBA diving and snorkel surveys to record rockfish densities associated with seagrass meadow habitat complexity, prey biomass, and predators.

Rockfish recruitment—YOY copper-quillback rockfish densities were quantified during three observation periods to establish seasonal recruitment dynamics: early (May–June), middle (early July), and late (July–August) summer. In nearshore habitats, rockfishes typically exhibit ontogenetic habitat shifts from an initial preference for the water column to later movement to benthic habitats (Carr 1991). To survey the full water column, both snorkel (high water column) and SCUBA (low water column) were used to record fish abundances and sizes on a

 40×4 m permanent transect at each site. All fish species were recorded on surveys and fish sizes were estimated using an underwater ruler. Rockfish 3-6 cm in length were classified as YOY recruits (Love et al. 1991). Observed current was recorded on a six-point categorical scale from none to very strong. In each observation period, sites were surveyed by SCUBA (three or four surveys) and snorkel (three or four surveys); site surveys were separated by several days. Each survey included an instant repeat transect pass by a different recorder to capture observer effects. Seagrass edge sites were surveyed 2 m into the meadow from the perimeter. Fish surveys were also conducted in N. luetkeana kelp forest and sand sites. A total of 791 surveys were conducted, hierarchically nested by observation period, site visit, visual method, and observer.

Seagrass habitat complexity and prey provision—During observation periods, metrics of seagrass habitat complexity and prey availability were collected at seagrass meadow sites. Using SCUBA, shoot density (count per 25×25 cm quadrat) and canopy height (maximum blade length) were measured every 5 m along the transect (n = 9 quadrats per transect). Divers collected seagrass shoots every 10 m along the transect (n = 5 shoots per transect) for shoot, epifauna, and epiphyte biomass. Shoots were covered with a collection bag and detached from the rhizome to ensure seagrass invertebrates (herein, "mesograzers") were captured in association with the shoot sample. Shoots were scraped of epiphytes and weighed (g). Length was measured from the first internode to blade tip. Scraped material was combined with free-floating material from the sample bag and passed through a 500-µm sieve prior to enumeration. Epifaunal mesograzers (>500 µm) were identified, counted, and weighed (mg). Epiphytic algae >500 µm (e.g., S. naiadum, Ulva sp., and Punctaria sp.) were also sorted and weighed (mg). Biomass samples (shoots, mesograzers, epiphytes) were dried at 60°C for 48 h to obtain dry mass.

Factors influencing YOY recruitment—We used generalized linear mixed-effects models (GLMMs) to determine the relative roles of seagrass habitat complexity, seagrass prey, and adjacent habitat type on YOY rockfish densities. To aggregate key properties of seagrass habitat complexity (biomass, density, and canopy height), a principal component analysis (PCA) was performed to obtain a singular seagrass habitat complexity metric (Appendix S3: Table S1). We used gammarid amphipod biomass to represent seagrass meadow prey provision because they were prevalent in both the YOY stomach contents and seagrass shoot collections. Black rockfish (S. maliger, >10 cm), copper rockfish (>10 cm), and kelp greenling (Hexagrammos decagrammus) densities were used to represent predators (Hobson et al. 2001). To account for collinearity between seagrass habitat complexity and other faunal densities (rockfish prey and

predators), we used the residual values of gammarid biomass and residual predator density from sequential linear regressions with seagrass habitat complexity (slope b = 0.19, P < 0.001, b = 0.12, P < 0.001; respectively) as predictors in the GLMMS (Graham 2003). An a priori model set was used to assess the strength of evidence for seagrass habitat complexity, seagrass prey, predators, and adjacent habitat type in predicting YOY recruitment (Appendix S3: Table S2). Candidate models varied in fixed effects, but included the same random effects: depth (chart datum), current, and diver nested within ordinal date, the latter to account for repeated surveys on the same day and over time. Continuous variables were standardized as per Gelman (2008). Model fits were compared using Akaike's information criterion corrected for small sample size (AIC_c) values, where the lowest relative AIC_c (ΔAIC_c) and highest relative weight (W_i) indicated the best model (Burnham and Anderson 2004).

RESULTS

Relative contribution of basal sources to YOY copperquillback rockfish

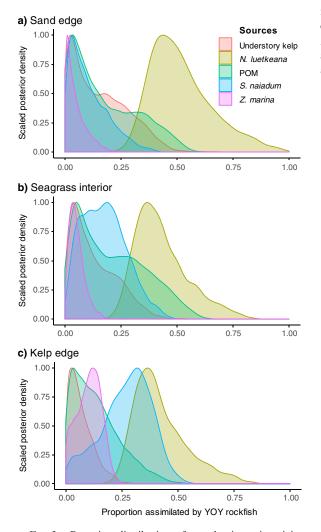
Allochthonous production was the dominant source of energy assimilated by YOY copper-quillback rockfish at all seagrass meadow sites, regardless of adjacent habitat (Fig. 3a–c; Appendix S2: Table S2). In particular, *N. luetkeana* macroalgae was the main source fueling YOY rockfish, contributing an average of $46\% \pm 7\%$ (mean \pm SD) of assimilated energy across all seagrass sites ($2.4\times$ greater than POM, the next highest contributor). The epiphyte *S. naiadum* was the largest contributor of autochthonous energy in YOY rockfish ($18\% \pm 6\%$). Contributions by *Z. marina* and understory kelp were low across sites.

The relative contribution of basal sources to YOY rockfish varied among sites (Fig. 3a-c). For YOY rockfish at sand edge sites, N. luetkeana macroalgae contributions were highest (53% \pm 14%), while autochthonous sources (e.g., S. naiadum and Z. marina) were minimal (Fig. 3a). In contrast, source contributions to YOY rockfish at kelp edge sites were more even (Fig. 3c). Autochthonous contributions were highest at these kelp edge sites, such that S. naiadum was the seccontributor ond highest $(27\% \pm 11\%),$ after N. luetkeana (42% \pm 12%). Similarly, S. naiadum was notable in the seagrass interior sites, albeit low overall $(18\% \pm 10\%, Fig. 3b).$

YOY rockfish diets and body condition

YOY copper-quillback rockfish diets demonstrated differences in prey composition across seagrass meadow sites (Fig. 4; Appendix S1: Fig. S1). At kelp edge sites, pelagically associated calanoid copepods were the dominant prey item in diets (47%), followed by harpacticoids

(16%). Conversely, YOY rockfish diets at sand edge sites primarily consisted of harpacticoid copepods (55%) and gammarid amphipods (28%). At seagrass interior sites, YOY rockfish exhibited intermediate diets of the edge sites (Fig. 4a), as well as being the most diverse and even (H' = 1.65, J = 0.85, respectively). However, the stomach sample size at the seagrass interior was low (n = 4, due to sample loss) relative to edge sites (n > 20) and may be a less reliable representation of prey composition at these sites. The body condition of YOY rockfish also varied across the seagrass



meadow, where YOY copper-quillback rockfish had higher body conditions at kelp edge and seagrass interior sites (Fig. 4b).

Patterns and drivers of rockfish recruitment in the seagrass seascape

YOY copper-quillback rockfish recruitment increased throughout the study period in Choked Passage, peaking in August, the last observation period (Fig. 5a). The highest recruitment in the seagrass meadow occurred consistently at the kelp edge sites, which was $6.3 \times$ and $2.2 \times$ greater than sand edge sites and seagrass interior sites, respectively. Overall, densities of YOY copper-quillback rockfish were greatest in adjacent *N. luetkeana* kelp forest sites, and consistently lowest in adjacent sand sites. Similarly, predators were most abundant in *N. luetkeana* kelp forest sites, followed by

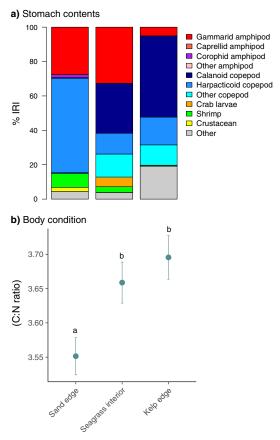


FIG. 3. Posterior distributions from the isotopic mixing model showing the relative proportional contribution of sources assimilated by YOY copper-quillback rockfish at sites in the seagrass meadow: (a) sand edge (n = 25 rockfish, n = 3 sites pooled), (b) seagrass interior (n = 29, n = 4 sites pooled), and (c) kelp edge sites (n = 32, n = 3 sites pooled). Sources in the model included understory kelps (*Cymathere triplicata* and *Alaria marginata*, red), *Nereocystis luetkeana* (light green), particulate organic matter (POM, dark green), *Zostera marina* (purple), and the epiphytic macroalgae *Smithora naiadum* (blue).

FIG. 4. (a) Prey consumed by YOY copper-quillback rockfish (index of relative importance, IRI) at sand edge (n = 29rockfish), seagrass interior (n = 4), and kelp edge (n = 19) sites. (b) YOY copper-quillback rockfish body condition (mean C:N ratio \pm SE) at sand edge (n = 25), seagrass interior (n = 29), and kelp edge (n = 32) sites. Means with the same letter are not significantly different from each other. See Appendix S1: Fig. S1 for numeric, gravimetric, and frequency of occurrence comparisons of prey.

kelp edge sites, where mean densities were significantly greater relative to seagrass interior and sand edge sites (Fig. 5b).

Multiple factors related to seascape connectivity influenced YOY copper-quillback rockfish densities in the seagrass meadow (Fig. 6; Appendix S3: Table S3). In the model with the highest support ($W_i = 1, \Delta AIC_c = 0$), seagrass habitat complexity and seagrass prey had positive effects on YOY copper-quillback rockfish recruitment in the meadow. Seascape features (i.e., edge and adjacent habitat type) had varying effects: (1) kelp edge sites had a large, positive effect on YOY densities, (2) sand edge sites had no evidence of an effect, and (3) interactions between seagrass habitat complexity and adjacent habitats influenced recruitment, such that adjacent habitats dampened

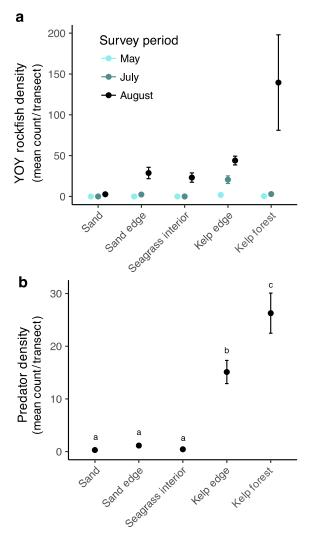


FIG. 5. (a) YOY copper-quillback rockfish densities (mean \pm SE) observed across sites and time. (b) Predator densities (mean \pm SE) pooled across the study period. Significant site differences among predator densities are distinguished by letters, while differences in YOY densities were reserved for model testing (Fig. 6).

the positive effects of seagrass habitat complexity. In addition, predators had a small negative effect on seagrass recruitment densities. Similar to YOY recruits, predators also had a positive association with seagrass complexity (b = 0.12, P < 0.001).

DISCUSSION

Seascape connectivity is prevalent in the nearshore, affecting the community structure of seagrass meadows (Staveley et al. 2017, Perry et al. 2018). Yet, mechanistic links to nursery function are lacking. We found multiple lines of evidence that connectivity to highly productive and structurally complex N. luetkeana forests enhanced Z. marina meadow nursery function for YOY rockfishes. First, N. luetkeana production fueled YOY copper-quillback rockfish in the seagrass meadow via allochthonous nutrient subsidies. Secondly, rockfish prey composition shifted between seagrass edge sites adjacent to kelp forests vs. unvegetated sand, with corresponding differences in body condition of YOY rockfish recruits (e.g., relatively higher lipids when adjacent to N. luetkeana kelp forests). Overall, the largest YOY recruitment in the seagrass meadow was observed adjacent to N. luetkeana kelp forests. Adjacent sand habitats, on the other hand, had little effect on YOY rockfish in the seagrass meadows. In sum, neighboring kelp forests within the seascape played a key role in the nursery effects of the seagrass meadow to YOY rockfish by increasing access to areas of greater habitat quality, as well as allochthonous subsidies.

Spatially mediated bottom-up subsidies from kelp forests

Our results reinforce the role of kelp forests as an important and pervasive allochthonous source of energy to nearshore marine food webs. Benthic macroalgae can become available to seagrass food webs by recruiting and growing within the same substrate used by seagrasses and/or, after detachment from nearby rocky substrates, accumulating in meadows with currents and tides (Hyndes et al. 2012, 2014). With higher nutritional content than seagrass leaves and high rates of consumption by grazers (Thayer et al. 1984, Smit et al. 2006, Doropoulos et al. 2009), macroalgae in general has the potential to make substantial contributions to seagrass food webs. Previous feeding trials (Doropoulos et al. 2009) and isotopic studies (Smit et al. 2006) have demonstrated a preference for macroalgae over seagrass by grazers, but could not distinguish between allochthonous or autochthonous (i.e., epiphytic) macroalgae. With distinct macroalgae isotopic signatures in this system (Appendix S2: Table S1), we were able to discern these differences and their relative contributions: N. luetkeana from adjacent kelp forests contributed more than epiphytic S. naiadum (found at high biomass within the seagrass meadow) and understory kelps (found at low biomass within the meadow).

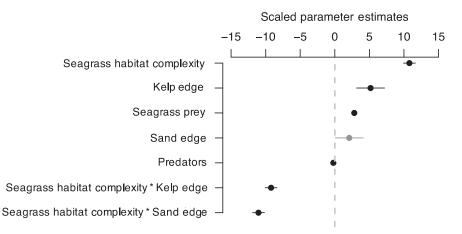


FIG. 6. Scaled parameter estimates (\pm SE) of seagrass nursery factors (seagrass habitat complexity and seagrass prey) and seascape factors (kelp edge, sand edge, and predator density) influencing YOY copper-quillback rockfish densities within the seagrass meadow. Significant parameter estimates (black) have error bars that do not cross zero (gray dashed line).

High contributions by N. luetkeana suggest that this allochthonous macroalgae is preferred by YOY rockfish prey (i.e., invertebrate grazers) over autochthonous sources, and this energy is transferred to top consumers. Differences in grazer preference for macroalgae species and traits are prevalent, and can depend on the chemical defenses, palatability, and/or morphology of macroalgae (Leighton 1966, Duffy and Hay 1994, Cox and Murray 2005). By utilizing N. luetkeana subsidies, seagrass meadow grazers may use epiphytic S. naiadum as habitat rather than food, as this behavior has been seen in low-mobility grazers (Duffy and Hay 1991) or in those under threat of predation (Boström and Mattila 1999). Preference for N. luetkeana may also be a function of its widespread distribution and availability in temperate nearshore environments (Kaehler et al. 2006, Krumhansl and Scheibling 2012, Ramshaw et al. 2017), but feeding trials are needed to test this hypothesis.

While N. luetkeana energy fueled YOY rockfish in the seagrass meadow, small-scale differences in the assimilation of N. luetkeana were nevertheless observed across sites. YOY rockfish diets adjacent to unvegetated sand habitats (i.e., sand edge sites) had the highest contribution of N. luetkeana subsidies. Although seagrass meadows are highly productive ecosystems, productivity can vary within a single meadow due to seascape configuration and environmental conditions (Fourgurean et al. 2001), which may lead to disproportionate effects of subsidies. Moreover, unvegetated sand shorelines tend to accumulate high rates of detrital macroalgae from adjacent reefs leading to important subsidies for invertebrate grazers in these habitats (Hyndes et al. 2014). As in other studies, areas of lower production and/or accumulation of drift macroalgae may benefit most from broadly available allochthonous kelp subsidies (Lastra et al. 2008, Gonçalves and Marques 2011, Kelly et al. 2012).

In addition to the provision of primary production, our results provide empirical support for spatially mediated trophic dynamics in seagrass meadows. Here, the structural complexity of adjacent habitats (high-complexity kelp forests vs. low-complexity unvegetated sand) altered prey availability and/or consumption in seagrass meadows. Small, juvenile fish in seagrass meadows have been found to select areas of high habitat complexity for shelter from predators without a compromise in foraging (Yeager and Hovel 2017). Increased habitat complexity in the water column with N. luetkeana forests (i.e., added vertical structure) may facilitate access to better prey. The prey of YOY rockfish adjacent to N. luetkeana forests was of higher quality than adjacent to sand, due to the dominance of lipid-rich, pelagic-associated calanoid copepods (Sargent and Falk-Petersen 1988) relative to bottom-associated harpacticoid copepods (Gee 1989), respectively. This dietary contrast indicates different feeding behavior (e.g., more pelagic) with increased habitat complexity and/or an alteration in zooplankton availability by the kelp forest (Pakhomov et al. 2002, Atilla et al. 2005).

Patterns in fish recruitment and trade-offs with habitat complexity

Three-dimensional habitat structure is a critical attribute promoting the nursery function of seagrasses (Heck et al. 2003, McDevitt-Irwin et al. 2016). We found that seagrass habitat complexity had the largest, positive effect on YOY rockfish recruitment density ($\sim 4 \times$ larger than prey provision). Yet, connectivity to structurally complex adjacent habitats can provide an additional enhancement of fish densities (Baltz et al. 1993, Dorenbosch et al. 2005, Valentine-Rose et al. 2007, Gilby et al. 2018). Adjacency to kelp forests had a large positive effect on YOY rockfish recruitment. Because habitat complexity is a key driver of juvenile recruitment success, edges occurring between two vegetated habitats may be optimal nursery areas in the seascape. However, habitat adjacency may also come with trade-offs such as predation. Because elevated predator densities were associated with both *N. luetkeana* and high seagrass habitat complexity (Figs. 5, 6), predator spillover from neighboring habitats is likely (Johnson 2006, Casini et al. 2012).

Access to multiple habitats is essential for many species during adult life stages and during ontogenetic shifts (Gillanders et al. 2003, Rand et al. 2006, Sheaves 2009). Incentives exist for predators to move into intermediate areas of habitat complexity (e.g., seagrass and kelp forest ecotones) to increase foraging success (Crowder and Cooper 1982, Johnson 2007). There, predators can cause increased mortality and/or shifts in habitat use by juveniles due to predation risk (Holbrook and Schmitt 1988, Persson and Eklov 1995), thereby dampening nursery benefits in these regions despite augmented refuge and subsidies. In this study, despite elevated predation risk at seagrass meadow edges, we found that other components of seascape connectivity largely controlled the observed rockfish nursery effects. Considering seascape composition and arrangement can reveal important trade-offs between optimizing foraging habitat and reducing predation (Lima and Dill 1990), factors that ultimately shape nursery function (Fig. 1).

Planning for seascape connectivity

Globally, human activity is accelerating the loss of seascape connections by fragmenting and altering habitats in nearshore areas. In British Columbia, anthropogenic disturbances are degrading seagrass habitats, in which rockfish species may be particularly sensitive (Iacarella et al. 2018). To counter these trends, our findings support the inclusion of a mosaic of nearshore habitats in protection zones where preserving and/or elevating nursery function is a primary goal of coastal conservation or management.

Based on this study, the ideal habitat mix should include habitat types that (1) increase structural complexity, fostering maximum access to resources and refugia and (2) augment productivity by increasing flows of production across habitat edges and seascape scales. Planning for habitat types that incorporate these components should also include an understanding of their temporal dynamism. For instance, kelp forests are strongly influenced by regional drivers (Krumhansl et al. 2016), where human-induced activities (removal of top predators, climate change, marine diseases) can influence declines in biomass (Watson and Estes 2011, Filbee-Dexter et al. 2016, Burt et al. 2018). As such, we advocate for the inclusion of a portfolio of habitat types when planning for seascape connectivity, including the incorporation of low complexity substrates capable of augmenting water column structural complexity in the future (e.g., rocky reef urchin barrens transitioning to

kelp forests), as well as an understanding that future seascape composition may change as a result of habitat loss.

Our findings also indicate that the arrangement of habitats within a seascape matters, particularly as a result of edge effects and composition of adjacent habitats. In this study, edge effects promoted access and use of seagrass edges, with trade-offs between predation, refuge, and provision of food. Overall, the observed edge effects had positive influences on nursery function for rockfish species, but varied by adjacent habitat type. Edge effects are ubiquitous across landscapes and seascapes, but can foster both positive and negative effects, depending on the ecological context and mechanisms at play (Ries et al. 2004). This study focused on the nursery function of a large, continuous seagrass meadow. But, nursery function may vary when habitat patches are smaller or seagrass meadows are fragmented, resulting in larger relative edge areas. Smaller seagrass patch sizes may illuminate different trade-offs with predation and subsidies (Smith et al. 2010). Moreover, fragmentation and/or habitat loss (Valentine-Rose et al. 2007), eutrophication (Deegan et al. 2002), and infrastructure additions (Bulleri and Chapman 2010), may further degrade the structure and function of coastal nursery habitats for juvenile fishes. As such, future research should also investigate the response of nursery function in areas of higher disturbance or in smaller seagrass meadows, which are often the management focus of coastal restoration or habitat mitigation.

The general application of these findings also depends on local and regional environmental factors influencing spatial connectivity, as well as the life history and traits of the focal fauna. Although not the focus of this study, tidal effects (Rangeley and Kramer 1995, 1998), upwelling dynamics (Bjorkstedt et al. 2002), and turbidity (Blaber and Blaber 1980) can alter juvenile habitat use and influence the physical connectivity between seascape habitats.

Fish species may also respond differently to local and regional environmental forcing, which may differentially affect recruitment, foraging behaviors or predatory avoidance strategies (Beukers and Jones 1998, Hammerschlag et al. 2010). While this study focused on rockfishes, more research is needed to investigate seascape nurseries across a wider range of fish species.

In general, seascape connectivity has been found to improve the efficacy of marine reserves (Halpern et al. 2010, Olds et al. 2011, 2015) and can increase the resilience of marine ecosystem functions (Mumby and Hastings 2007, Olds et al. 2012). For temperate coastal seascapes, where connectivity is most at risk by habitat loss and climate change (Hyndes et al. 2014), our findings illustrate how the inclusion of habitat mosaics, and connectivity between habitats, can be utilized advantageously by marine planners and managers to promote fish nurseries. In BC, Rockfish Conservation Areas (RCAs) have been established coast-wide to protect declining rockfish populations (Yamanaka and Logan 2010, Haggarty et al. 2016). Although a long recovery time for rockfish populations is expected due to their life-history characteristics, overall positive reserve effects have not yet been found and RCAs are being evaluated for their effectiveness (Haggarty et al. 2016). Evaluation of habitat in and around RCAs (Haggarty and Yama-naka 2018) should consider connectivity to seascape nurseries, including seagrass meadows, kelp forests, and high-relief rocky reefs to improve RCA effectiveness and the recovery of rockfish abundances.

Our research also emphasizes the importance of seascape connectivity in a large-scale Marine Protected Area (MPA) Network planning processes such as the one ongoing in Canada's Northern Shelf Bioregion, which aims to protect 10% of its coastal and marine areas by 2020 (DFO 2017). Although kelp and seagrass habitats are considered Ecologically and Biologically Significant Areas (EBSAs; Rubidge et al. 2018) and receive conservation priorities within MPA network design (DFO 2017), little empirical evidence currently informs design guidelines aimed at promoting connectivity among these habitats. Our research provides support for the prioritization of connected seascape habitats in these protected area networks.

CONCLUSION

Concurrent with emerging evidence (Nagelkerken et al. 2013, Perry et al. 2018), our findings highlight the importance of temperate seascape connectivity in positively influencing seagrass nursery function. We recommend the protection and management of nearshore habitats mosaics to maintain and/or enhance nursery function and value of coastal ecosystems for commercially and biologically important fish populations.

ACKNOWLEDGMENTS

The Tula Foundation supported this research through a Hakai Scholarship to A. M. Olson and support to M. Hessing-Lewis. Additional support came from the UVic Biology Department to A. M. Olson, and from NSERC, CFI/BCKDF, and the Liber Ero Foundation to F. Juanes. Thank you to all who assisted in the field, particularly C. Prentice, D. VanMaanen, and T. Blaine. Thank you to Juanes Lab members: G. Sadlier-Brown, L. Simon, and L. Campbell for microscope work. The Hakai Institute provided the seascape map and meadow size estimation (L. Reshitnyk), and POM isotope values (B. P. V. Hunt). We are grateful to B. Hughes, A. K. Salomon, Z. Monteith, and E. Rechsteiner for input into various stages of this project. We acknowledge the use of graphics by J. Silberg (Hakai), T. Saxby, and D. Tracey (IAN, University of Maryland, Center for Environmental Science, ian.umces.edu/ imagelibrary/).

LITERATURE CITED

Atilla, N., J. W. Fleeger, and C. M. Finelli. 2005. Effects of habitat complexity and hydrodynamics on the abundance and diversity of small invertebrates colonizing artificial substrates. Journal of Marine Research 63:1151–1172.

- Burnham, K. P., and D. R. Anderson. 2004. Multimodel inference: understanding AIC and BIC in model selection. Sociological methods & research 33:261–304.
- Baltz, D. M., C. Rakocinski, and J. W. Fleeger. 1993. Microhabitat use by marsh-edge fishes in a Louisiana estuary. Environmental Biology of Fishes 36:109–126.
- Beck, M. W., et al. 2001. The identification, conservation, and management of estuarine and marine nurseries for fish and invertebrates. BioScience 51:633.
- Berkström, C., R. Lindborg, M. Thyresson, and M. Gullström. 2013. Assessing connectivity in a tropical embayment: fish migrations and seascape ecology. Biological Conservation 166:43–53.
- Beukers, J. S., and G. P. Jones. 1998. Habitat complexity modifies the impact of piscivores on a coral reef fish population. Oecologia 114:50–59.
- Bjorkstedt, E. P., L. K. Rosenfeld, B. A. Grantham, Y. Shkedy, and J. Roughgarden. 2002. Distributions of larval rockfishes *Sebastes* spp. across nearshore fronts in a coastal upwelling region. Marine Ecology Progress Series 242:215–228.
- Blaber, S. J. M., and T. G. Blaber. 1980. Factors affecting the distribution of juvenile estuarine and inshore fish. Journal of Fish Biology 17:143–162.
- Bond, A. L., and A. W. Diamond. 2011. Recent Bayesian stable-isotope mixing models are highly sensitive to variation in discrimination factors. Ecological Applications 21:1017– 1023.
- Boström, C., E. L. Jackson, and C. A. Simenstad. 2006. Seagrass landscapes and their effects on associated fauna: a review. Estuarine, Coastal and Shelf Science 68:383–403.
- Boström, C., and J. Mattila. 1999. The relative importance of food and shelter for seagrass-associated invertebrates: a latitudinal comparison of habitat choice by isopod grazers. Oecologia 120:162–170.
- Boström, C., S. J. Pittman, C. Simenstad, and R. T. Kneib. 2011. Seascape ecology of coastal biogenic habitats: advances, gaps, and challenges. Marine Ecology Progress Series 427:191–217.
- Bulleri, F., and M. G. Chapman. 2010. The introduction of coastal infrastructure as a driver of change in marine environments. Journal of Applied Ecology 47:26–35.
- Burt, J. M., M. T. Tinker, D. K. Okamoto, K. W. Demes, K. Holmes, and A. K. Salomon. 2018. Sudden collapse of a mesopredator reveals its complementary role in mediating rocky reef regime shifts. Proceedings of the Royal Society B 285:20180553.
- Carr, M. H. 1991. Habitat selection and recruitment of an assemblage of temperate zone reef fishes. Journal of Experimental Marine Biology and Ecology 146:113–137.
- Casini, M., T. Blenckner, C. Möllmann, A. Gårdmark, M. Lindegren, M. Llope, G. Kornilovs, M. Plikshs, and N. C. Stenseth. 2012. Predator transitory spillover induces trophic cascades in ecological sinks. Proceedings of the National Academy of Sciences USA 109:8185–8189.
- Connell, S. D., and G. P. Jones. 1991. The influence of habitat complexity on postrecruitment processes in a temperate reef fish population. Journal of Experimental Marine Biology and Ecology 151:271–294.
- Cox, T. E., and S. N. Murray. 2005. Feeding preferences and the relationships between food choice and assimilation efficiency in the herbivorous marine snail *Lithopoma undosum* (Turbinidae). Marine Biology 148:1295–1306.
- Crowder, L. B., and W. E. Cooper. 1982. Habitat structural complexity and the interaction between bluegills and their prey. Ecology 63:1802.
- Dale, A. W., and R. Prego. 2002. Physico-biogeochemical controls on benthic-pelagic coupling of nutrient fluxes and

recycling in a coastal upwelling system. Marine Ecology Progress Series 235:15–28.

- Davis, J. P., K. A. Pitt, B. Fry, A. D. Olds, and R. M. Connolly. 2014. Seascape-scale trophic links for fish on inshore coral reefs. Coral Reefs 33:897–907.
- Deegan, L. A., A. Wright, S. G. Ayvazian, J. T. Finn, H. Golden, R. R. Merson, and J. Harrison. 2002. Nitrogen loading alters seagrass ecosystem structure and support of higher trophic levels. Aquatic Conservation: Marine and Freshwater Ecosystems 12:193–212.
- Deniro, M. J., and S. Epstein. 1981. Influence of diet on the distribution of nitrogen isotopes in animals. Geochimica et Cosmochimica Acta 45:341–351.
- DFO. 2017. Framework for identification of ecological conservation priorities for marine protected area network design and its application in the northern shelf bioregion. DFO Canadian Science Advisory Secretariat Science Advisory Report. Fisheries and Oceans Canada. Ottawa, Canada. 2017/019 (Errata: October 2018).
- Dorenbosch, M., M. Grol, M. Christianen, I. Nagelkerken, and G. van der Velde. 2005. Indo-Pacific seagrass beds and mangroves contribute to fish density and diversity on adjacent coral reefs. Marine Ecology Progress Series 302:63–76.
- Doropoulos, C., G. A. Hyndes, P. S. Lavery, and F. Tuya. 2009. Dietary preferences of two seagrass inhabiting gastropods: Allochthonous vs. autochthonous resources. Estuarine, Coastal and Shelf Science 83:13–18.
- Duffy, J. E., and M. E. Hay. 1991. Food and shelter as determinants of food choice by an herbivorous marine amphipod. Ecology 72:1286–1298.
- Duffy, J. E., and M. E. Hay. 1994. Herbivore resistance to seaweed chemical defense: the roles of mobility and predation risk. Ecology 75:1304–1319.
- Eckert, L. E., N. C. Ban, A. Frid, and M. McGreer. 2017. Diving back in time: extending historical baselines for yelloweye rockfish with Indigenous knowledge. Aquatic Conservation: Marine and Freshwater Ecosystems 28:158–166.
- Ellison, A. M., et al. 2005. Loss of foundation species: consequences for the structure and dynamics of forested ecosystems. Frontiers in Ecology and the Environment 3:479–486.
- Fagan, W. F., R. S. Cantrell, and C. Cosner. 1999. How habitat edges change species interactions. American Naturalist 153:165–182.
- Filbee-Dexter, K., C. J. Feehan, and R. E. Scheibling. 2016. Large-scale degradation of a kelp ecosystem in an ocean warming hotspot. Marine Ecology Progress Series 543:141– 152.
- Fourqurean, J. W., A. Willsie, C. D. Rose, and L. M. Rutten. 2001. Spatial and temporal pattern in seagrass community composition and productivity in south Florida. Marine Biology 138:341–354.
- Gee, J. M. 1989. An ecological and economic review of meiofauna as food for fish. Zoological Journal of the Linnean Society 96:243–261.
- Gelman, A. 2008. Scaling regression inputs by dividing by two standard deviations. Statistics in Medicine 27:2865–2873.
- Gilby, B. L., A. D. Olds, R. M. Connolly, P. S. Maxwell, C. J. Henderson, and T. A. Schlacher. 2018. Seagrass meadows shape fish assemblages across estuarine seascapes. Marine Ecology Progress Series 588:179–189.
- Gillanders, B. M., K. W. Able, J. A. Brown, D. B. Eggleston, and P. F. Sheridan. 2003. Evidence of connectivity between juvenile and adult habitats for mobile marine fauna: an important component of nurseries. Marine Ecology Progress Series 247:281–295.
- Gonçalves, S. C., and J. C. Marques. 2011. The effects of season and wrack subsidy on the community functioning of exposed

sandy beaches. Estuarine, Coastal and Shelf Science 95:165-177.

- Graham, M. H. 2003. Confronting multicollinearity in ecological multiple regression. Ecology 84:2809–2815.
- Graham, M. H. 2004. Effects of local deforestation on the diversity and structure of Southern California giant kelp forest food webs. Ecosystems 7:341–357.
- Grober-Dunsmore, R., S. J. Pittman, C. Caldow, M. S. Kendall, and T. K. Frazer. 2009. A landscape ecology approach for the study of ecological connectivity across tropical marine seascapes. Pages 493–530 *in* Ecological connectivity among tropical coastal ecosystems. Springer, Dordrecht, The Netherlands.
- Haddad, N. M., et al. 2015. Habitat fragmentation and its lasting impact on Earth's ecosystems. Science Advances 1: e1500052.
- Haggarty, D. R., K. E. Lotterhos, and J. B. Shurin. 2017. Young-of-the-year recruitment does not predict the abundance of older age classes in Black rockfish in Barkley Sound, British Columbia, Canada. Marine Ecology Progress Series 574:113–126.
- Haggarty, D. R., J. B. Shurin, and L. K. Yamanaka. 2016. Assessing population recovery inside British Columbia's Rockfish Conservation Areas with a remotely operated vehicle. Fisheries Research 183:165–179.
- Haggarty, D. R., and L. K. Yamanaka. 2018. Evaluating Rockfish Conservation Areas in southern British Columbia, Canada using a Random Forest model of rocky reef habitat. Estuarine, Coastal and Shelf Science 208:191–204.
- Halpern, B. S., S. E. Lester, and K. L. McLeod. 2010. Placing marine protected areas onto the ecosystem-based management seascape. Proceedings of the National Academy of Sciences USA 107:18312–18317.
- Hammerschlag, N., M. R. Heithaus, and J. E. Serafy. 2010. Influence of predation risk and food supply on nocturnal fish foraging distributions along a mangrove–seagrass ecotone. Marine Ecology Progress Series 414:223–235.
- Hart, R. K., M. C. Calver, and C. R. Dickman. 2002. The index of relative importance: an alternative approach to reducing bias in descriptive studies of animal diets. Wildlife Research 29:415.
- Heck Jr., K. L., G. Hays, and R. J. Orth. 2003. Critical evaluation of the nursery role hypothesis for seagrass meadows. Marine Ecology Progress Series 253:123–136.
- Hessing-Lewis, M., E. U. Rechsteiner, B. B. Hughes, M. T. Tinker, Z. L. Monteith, A. M. Olson, M. M. Henderson, and J. C. Watson. 2018. Ecosystem features determine seagrass community response to sea otter foraging. Marine Pollution Bulletin 134:134–144.
- Hoelzer, G. A. 1988. Juvenile movement patterns in a territorial scorpaenid fish before and during settlement. Marine Ecology Progress Series 45:193–195.
- Holbrook, S. J., and R. J. Schmitt. 1988. The combined effects of predation risk and food reward on patch selection. Ecology 69:125–134.
- Hobson, E. S., R. C. James, and D. F. Howard. 2001. Interannual variation in predation on first-year *Sebastes* spp. by three northern California predators. Fishery Bulletin 99:292.
- Hughes, B. B., R. Eby, E. Van Dyke, M. T. Tinker, C. I. Marks, K. S. Johnson, and K. Wasson. 2013. Recovery of a top predator mediates negative eutrophic effects on seagrass. Proceedings of the National Academy of Sciences USA 110:15313–15318.
- Hughes, A. R., S. L. Williams, C. M. Duarte, K. L. Heck Jr., and M. Waycott. 2009. Associations of concern: declining seagrasses and threatened dependent species. Frontiers in Ecology and the Environment 7:242–246.

- Hyndes, G. A., P. S. Lavery, and C. Doropoulos. 2012. Dual processes for cross-boundary subsidies: incorporation of nutrients from reef-derived kelp into a seagrass ecosystem. Marine Ecology Progress Series 445:97–107.
- Hyndes, G. A., I. Nagelkerken, R. J. McLeod, R. M. Connolly, P. S. Lavery, and M. A. Vanderklift. 2014. Mechanisms and ecological role of carbon transfer within coastal seascapes. Biological Reviews of the Cambridge Philosophical Society 89:232–254.
- Iacarella, J. C., et al. 2018. Anthropogenic disturbance homogenizes seagrass fish communities. Global Change Biology 24:1904–1918.
- Johnson, D. W. 2006. Predation, habitat complexity, and variation in density-dependent mortality of temperate reef fishes. Ecology 87:1179–1188.
- Johnson, D. W. 2007. Habitat complexity modifies post-settlement mortality and recruitment dynamics of a marine fish. Ecology 88:1716–1725.
- Juanes, F. 2007. Role of habitat in mediating mortality during the post-settlement transition phase of temperate marine fishes. Journal of Fish Biology 70:661–677.
- Kaehler, S., E. A. Pakhomov, R. M. Kalin, and S. Davis. 2006. Trophic importance of kelp-derived suspended particulate matter in a through-flow sub-Antarctic system. Marine Ecology Progress Series 316:17–22.
- Kamimura, Y., and J. Shoji. 2013. A combination of seagrass and macroalgal beds enhances survival of young-of-the-year rockfish *Sebastes cheni*. Pages 61–68 in E. Moksness, E. Dahl, and J. Støttrup, editors. Global challenges in integrated coastal zone management. John Wiley & Sons, Oxford, UK.
- Kelly, J. R., K. A. Krumhansl, and R. E. Scheibling. 2012. Drift algal subsidies to sea urchins in low-productivity habitats. Marine Ecology Progress Series 452:145–157.
- Kennedy, L. A., F. Juanes, and R. El-Sabaawi. 2018. Eelgrass as valuable nearshore foraging habitat for juvenile pacific salmon in the early marine period. Marine and Coastal Fisheries 10:190–203.
- Krumhansl, K. A., and R. E. Scheibling. 2012. Production and fate of kelp detritus. Marine Ecology Progress Series 467:281–302.
- Krumhansl, K. A., et al. 2016. Global patterns of kelp forest change over the past half-century. Proceedings of the National Academy of Sciences USA 113:13785–13790.
- Lastra, M., H. M. Page, J. E. Dugan, D. M. Hubbard, and I. F. Rodil. 2008. Processing of allochthonous macrophyte subsidies by sandy beach consumers: estimates of feeding rates and impacts on food resources. Marine Biology 154:163–174.
- Leighton, D. L. 1966. Studies of food preference in algivorous invertebrates of Southern California kelp beds. Pacific Science 20:104–113.
- Liao, H., C. L. Pierce, and J. G. Larscheid. 2001. Empirical assessment of indices of prey importance in the diets of predacious fish. Transactions of the American Fisheries Society 130:583–591.
- Lima, S. L., and L. M. Dill. 1990. Behavioral decisions made under the risk of predation: a review and prospectus. Canadian Journal of Zoology 68:619–640.
- Loreau, M., N. Mouquet, and R. D. Holt. 2003. Meta-ecosystems: a theoretical framework for a spatial ecosystem ecology. Ecology Letters 6:673–679.
- Love, M. S., M. H. Carr, and L. J. Haldorson. 1991. The ecology of substrate-associated juveniles of the genus *Sebastes*. Environmental Biology of Fishes 30:225–243.
- Love, M. S., M. Yoklavich, and L. K. Thorsteinson. 2002. The rockfishes of the Northeast Pacific. University of California Press, Berkeley, California, USA.

- Massol, F., D. Gravel, N. Mouquet, M. W. Cadotte, T. Fukami, and M. A. Leibold. 2011. Linking community and ecosystem dynamics through spatial ecology. Ecology Letters 14:313–323.
- McCutchan, J. H., W. M. Lewis, C. Kendall, and C. C. McGrath. 2003. Variation in trophic shift for stable isotope ratios of carbon, nitrogen, and sulfur. Oikos 102:378–390.
- McDevitt-Irwin, J. M., J. C. Iacarella, and J. K. Baum. 2016. Reassessing the nursery role of seagrass habitats from temperate to tropical regions: a meta-analysis. Marine Ecology Progress Series 557:133–143.
- McMahon, K. W., M. L. Berumen, and S. R. Thorrold. 2012. Linking habitat mosaics and connectivity in a coral reef seascape. Proceedings of the National Academy of Sciences USA 109:15372–15376.
- Minagawa, M., and E. Wada. 1984. Stepwise enrichment of ${}^{15}N$ along food chains: further evidence and the relation between $\delta^{15}N$ and animal age. Geochimica et Cosmochimica Acta 48:1135–1140.
- Moore, J. W., and B. X. Semmens. 2008. Incorporating uncertainty and prior information into stable isotope mixing models. Ecology Letters 11:470–480.
- Mumby, P. J., and A. Hastings. 2007. The impact of ecosystem connectivity on coral reef resilience. Journal of Applied Ecology 45:854–862.
- Nagelkerken, I., C. M. Roberts, G. van der Velde, M. Dorenbosch, M. C. van Riel, E. C. de la Morinière, and P. H. Nienhuis. 2002. How important are mangroves and seagrass beds for coral-reef fish? The nursery hypothesis tested on an island scale. Marine Ecology Progress Series 244:299–305.
- Nagelkerken, I., M. Sheaves, R. Baker, and R. M. Connolly. 2013. The seascape nursery: a novel spatial approach to identify and manage nurseries for coastal marine fauna. Fish and Fisheries 16:362–371.
- Naiman, R. J., R. E. Bilby, D. E. Schindler, and J. M. Helfield. 2002. Pacific salmon, nutrients, and the dynamics of freshwater and riparian ecosystems. Ecosystems 5:399–417.
- Nelson, P. A. 2001. Behavioral ecology of young-of-the-year kelp rockfish, *Sebastes atrovirens* Jordan and Gilbert (Pisces: Scorpaenidae). Journal of Experimental Marine Biology and Ecology 256:33–50.
- Norton, E. C., R. B. MacFarlane, and M. S. Mohr. 2001. Lipid class dynamics during development in early life stages of shortbelly rockfish and their application to condition assessment. Journal of Fish Biology 58:1010–1024.
- Olds, A. D., R. M. Connolly, K. A. Pitt, and P. S. Maxwell. 2011. Habitat connectivity improves reserve performance. Conservation Letters 5:56–63.
- Olds, A. D., R. M. Connolly, K. A. Pitt, and P. S. Maxwell. 2012. Primacy of seascape connectivity effects in structuring coral reef fish assemblages. Marine Ecology Progress Series 462:191–203.
- Olds, A. D., et al. 2015. Quantifying the conservation value of seascape connectivity: a global synthesis. Global Ecology and Biogeography 25:3–15.
- Orth, R. J., K. L. Heck, and J. van Montfrans. 1984. Faunal communities in seagrass beds: A review of the influence of plant structure and prey characteristics on predator: prey relationships. Estuaries 7:339–350.
- Orth, R. J., et al. 2006. A global crisis for seagrass ecosystems. BioScience 56:987–996.
- Pakhomov, E. A., S. Kaehler, and C. D. McQuaid. 2002. Zooplankton community structure in the kelp beds of the sub-Antarctic Prince Edward Archipelago: Are they a refuge for larval stages? Polar Biology 25:778–788.
- Perry, D., T. A. B. Staveley, and M. Gullström. 2018. Habitat connectivity of fish in temperate shallow-water seascapes. Frontiers in Marine Science 4:5–12.

- Persson, L., and P. Eklov. 1995. Prey refuges affecting interactions between piscivorous perch and juvenile perch and roach. Ecology 76:70–81.
- Phillips, D. L., R. Inger, S. Bearhop, A. L. Jackson, J. W. Moore, A. C. Parnell, B. X. Semmens, and E. J. Ward. 2014. Best practices for use of stable isotope mixing models in food-web studies. Canadian Journal of Zoology 92:823–835.
- Polis, G. A., W. B. Anderson, and R. D. Holt. 1997. Toward an integration of landscape and food web ecology: the dynamics of spatially subsidized food webs. Annual Review of Ecology and Systematics 28:289–316.
- Post, D. M., C. A. Layman, D. A. Arrington, G. Takimoto, J. Quattrochi, and C. G. Montaña. 2007. Getting to the fat of the matter: models, methods and assumptions for dealing with lipids in stable isotope analyses. Oecologia 152:179–189.
- Ramshaw, B. C., E. A. Pakhomov, R. W. Markel, and S. Kaehler. 2017. Quantifying spatial and temporal variations in phytoplankton and kelp isotopic signatures to estimate the distribution of kelp-derived detritus off the west coast of Vancouver Island, Canada. Limnology and Oceanography 62:2133–2153.
- Rand, T. A., J. M. Tylianakis, and T. Tscharntke. 2006. Spillover edge effects: the dispersal of agriculturally subsidized insect natural enemies into adjacent natural habitats. Ecology Letters 9:603–614.
- Rangeley, R. W., and D. L. Kramer. 1995. Tidal effects on habitat selection and aggregation by juvenile pollock *Pollachius virens* in the rocky intertidal zone. Marine Ecology Progress Series 126:19–29.
- Rangeley, R. W., and D. L. Kramer. 1998. Density-dependent antipredator tactics and habitat selection in juvenile pollock. Ecology 79:943–952.
- Reid, D. J., G. P. Quinn, P. S. Lake, and P. Reich. 2008. Terrestrial detritus supports the food webs in lowland intermittent streams of south-eastern Australia: a stable isotope study. Freshwater Biology 53:2036–2050.
- Ricart, A. M., A. Dalmau, M. Pérez, and J. Romero. 2015. Effects of landscape configuration on the exchange of materials in seagrass ecosystems. Marine Ecology Progress Series 532:89–100.
- Ries, L., R. J. Fletcher, J. Battin, and T. D. Sisk. 2004. Ecological responses to habitat edges: mechanisms, models, and variability explained. Annual Review of Ecology, Evolution, and Systematics 35:491–522.
- Rubidge, E., J. Nephin, K. S. P. Gale, and J. Curtis. 2018. Reassessment of the ecologically and biologically significant areas (EBSAs) in the Pacific Northern Shelf Bioregion. DFO Canadian Science Advisory Secretariat Research Document. Fisheries and Oceans Canada. Ottawa, Canada. 2018/053. Fisheries and Oceans Canada, Ottawa, Canada.

- Sargent, J. R., and S. Falk-Petersen. 1988. The lipid biochemistry of calanoid copepods. Hydrobiologia 167–168:101–114.
- Sheaves, M. 2009. Consequences of ecological connectivity: the coastal ecosystem mosaic. Marine Ecology Progress Series 391:107–115.
- Smit, A. J., A. Brearley, G. A. Hyndes, P. S. Lavery, and D. I. Walker. 2006. 815N and 813C analysis of a *Posidonia sinuosa* seagrass bed. Aquatic Botany 84:277–282.
- Smith, T. M., J. S. Hindell, G. P. Jenkins, and R. M. Connolly. 2010. Seagrass patch size affects fish responses to edges. Journal of Animal Ecology 79:275–281.
- Staveley, T. A. B., D. Perry, R. Lindborg, and M. Gullström. 2017. Seascape structure and complexity influence temperate seagrass fish assemblage composition. Ecography 40:936–946.
- Thayer, G. W., K. A. Bjorndal, J. C. Ogden, S. L. Williams, and J. C. Zieman. 1984. Role of larger herbivores in seagrass communities. Estuaries 7:351–376.
- Thompson, P. L., B. Rayfield, and A. Gonzalez. 2017. Loss of habitat and connectivity erodes species diversity, ecosystem functioning, and stability in metacommunity networks. Ecography 40:98–108.
- Trebilco, R., N. K. Dulvy, H. Stewart, and A. K. Salomon. 2015. The role of habitat complexity in shaping the size structure of a temperate reef fish community. Marine Ecology Progress Series 532:197–211.
- Tupper, M., and F. Juanes. 2017. Testing foraging arena theory: the effects of conspecific density and habitat type on time and energy budgets of juvenile cunner. Journal of Experimental Marine Biology and Ecology 487:86–93.
- Valentine-Rose, L., C. A. Layman, D. A. Arrington, and A. L. Rypel. 2007. Habitat fragmentation decreases fish secondary production in Bahamian tidal creeks. Bulletin of Marine Science 80:863–877.
- Van Wynsberge, S., S. Andréfouët, N. Gaertner-Mazouni, J. Tiavouane, D. Grulois, J. Lefèvre, M. L. Pinsky, and C. Fauvelot. 2017. Considering reefscape configuration and composition in biophysical models advance seascape genetics. PloS one 12(5):e0178239.
- Watson, J., and J. A. Estes. 2011. Stability, resilience, and phase shifts in rocky subtidal communities along the west coast of Vancouver Island, Canada. Ecological Monographs 81:215–239.
- Waycott, M., et al. 2009. Accelerating loss of seagrasses across the globe threatens coastal ecosystems. Proceedings of the National Academy of Sciences USA 106:12377–12381.
- Yamanaka, K. L., and G. Logan. 2010. Developing British Columbia's inshore rockfish conservation strategy. Marine and Coastal Fisheries 2:28–46.
- Yeager, M. E., and K. A. Hovel. 2017. Structural complexity and fish body size interactively affect habitat optimality. Oecologia 185:257–267.

SUPPORTING INFORMATION

Additional supporting information may be found online at: http://onlinelibrary.wiley.com/doi/10.1002/eap.1897/full

DATA AVAILABILITY

Data are available in the Hakai Metadata Catalogue: https://doi.org/10.21966/exy6-1k58